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The spectral species concept in living color

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Abstract

Biodiversity monitoring is an almost inconceivable challenge at the scale of the entire Earth. The current (and soon to be flown) generation of spaceborne and airborne optical sensors (i.e., imaging spectrometers) can collect detailed information at unprecedented spatial, temporal, and spectral resolutions. These new data streams are preceded by a revolution in modelling and analytics that can utilize the richness of these datasets to measure a wide range of plant traits, community composition, and ecosystem functions. At the heart of this framework for monitoring plant biodiversity is the idea of remotely identifying species by making use of the 'spectral species' concept. In theory, the spectral species concept can be defined as a species characterized by a unique spectral signature and thus remotely detectable within pixel units of a spectral image. In reality, depending on spatial resolution, pixels may contain several species which renders speciesspecific assignment of spectral information more challenging. The aim of this paper is to review the spectral species concept and relate it to underlying ecological principles, while also discussing the complexities, challenges and opportunities to apply this concept given current and future scientific advances in remote sensing.

Keywords: airborne sensors; biodiversity; ecoinformatics; hyperspectral
 images; plant optical types; remote sensing; satellite imagery; vegetation
 communities.

³⁵ Plain language summary (PLS)

Biodiversity monitoring based on field data in almost inconceivable at the 96 scale of the entire Earth. Over the past decades, remote sensing has opened 97 possibilities for Earth observation from air and space, allowing us to mon-98 itor ecological change, primarily expressed by changes in vegetation cover, 99 distribution and functioning, which can be subsequently linked to drivers of 100 change in space and time, from local to global scale. Recently, the spectral 101 species concept - an algorithm that clusterizes pixels from spectral images 102 with a similar spectral signal (referred to as 'spectral species') - has brought 103 attention. The aim of this paper is to review the ecological functioning 104 principles of the spectral species concept and to refine its definition by a 105 better linkage with field observations of plant species distribution data (i.e. 106 presence-absence data) available from vegetation surveys. 107

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108 Key Points (KP)

- 1. Remote sensing has opened possibilities for Earth observation from air and space, allowing us to monitor ecological change.
- 2. Biodiversity monitoring based on field data in almost inconceivable at the scale of the entire Earth.
- 3. The spectral species concept, relating field to remotely sensed data, can open news ways to measure diversity from space.

115 1 Background

Rapid environmental changes are occurring across the globe at small to 116 large spatial extents due to the combined effects of climate change, land-use 117 changes, and biological invasions (Kreft & Jetz, 2007). Therefore, there is 118 an urgent pressing demand for operational biodiversity monitoring systems 119 that will improve our understanding of the repercussions of these drivers 120 of changes on ecosystem functioning and lead to better ecosystem manage-121 ment (Skidmore et al., 2021). New approaches are required to obtain timely 122 biodiversity data that are consistently and routinely measured across the 123 Earth surface. Some of these needs are fulfilled by remote sensing informa-124 tion (Schweiger and Laliberté, 2022). Over the past decades, remote sensing 125 has opened possibilities for Earth observation from air and space, allowing 126 us to monitor ecological change, primarily expressed by changes in vegeta-127 tion cover, distribution and functioning, which can be subsequently linked 128 to drivers of change in space and time, from local to global scale (Skidmore 129 et al., 2015; Asner et al., 2017). Recent technological advances in remote 130 sensing data acquisition and processing now open new perspectives for moni-131 toring changes in biodiversity at unprecedented details over large geographic 132 areas, and ultimately over the entire Earth (Luque et al., 2018; Randin et al., 133 2020). Furthermore, missions like the Surface Biology and Geology (SBG) 134 by NASA (https://sbg.jpl.nasa.gov/, Cawse-Nicholson et al., 2021) have 135 been implemented to support the development of algorithms for exploiting 136 spaceborne remotely sensed data and providing a relatively fast but accurate 137 estimate of ecological properties in vast areas over time. 138

Spaceborne and airborne passive optical sensors relying on imaging spec-139 troscopy (i.e., spectral remote sensing including multispectral and hyperspec-140 tral imaging) are a good example of the recent remote-sensing revolution in 141 ecology (Kwok, 2018). By measuring information from most of the electro-142 magnetic spectrum operable for Earth observation, imaging spectroscopy has 143 demonstrated significant capabilities to detect and monitor the spatial dis-144 tribution of plant communities, species and traits (Asner & Martin, 2008; 145 Schaaf et al., 2013; Schweiger et al., 2017; Skowronek et al., 2017). The pixel 146 reflectance in an optical image results from the integration of multiple inter-147 actions between light and matter, including vegetation and the surrounding 148 environment (soil, atmosphere). Intrinsic properties of vegetation influenc-149 ing this remotely sensed information correspond to biophysical and biochem-150 ical properties (i.e., traits) of leaves and the canopy that can be related to 151 levels of ecological organizations such as ecosystems, communities, species, 152 and potentially to the intraspecific trait of plant genotypes (Madritch et al., 153 2014; Blonder et al., 2020). Spectroscopy has long been used to capture the 154

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characteristic absorption features of biochemical compounds of plants, which 155 biologically corresponds to the phenotypic expression of some of the genes 156 that describe individuals belonging to a given species (e.g., Jacquemoud & 157 Ustin, 2019). These biochemical traits and their dynamics are linked to 158 functional traits, opening access to the monitoring of ecosystem functions, 159 processes and services. Besides, imaging spectroscopy has already demon-160 strated capabilities for species discrimination in various types of ecosystems 161 (Féret & Asner, 2014; Fassnacht et al., 2016; Skowronek et al., 2017). At the 162 sub-organism level, biochemical properties estimated from observations at 163 leaf and canopy scales (Baret et al., 1994; Kokaly et al., 2009; Ollinger, 2011; 164 Serbin et al., 2012) are described in commonly used radiative transfer mod-165 els (Jacquemoud et al., 1996; Féret et al., 2008; Torresani et al., 2021). The 166 fine spectral resolution and spectral sampling interval on imaging spectrome-167 ters provide information to quantify key biochemical properties of vegetation 168 such as leaf pigment (e.g., chlorophylls, carotenoids, anthocyanins), water, 169 cellulose, lignin, nitrogen, phosphorous and protein contents based on their 170 specific light absorption characteristics (Ewald et al., 2018). 171

The link between biochemical properties, functional attributes (morpho-172 logical, physiological and phenological traits), and taxonomic information is 173 often implicitly assumed when performing spectroscopic analysis for biodiversity monitoring. Recently, the spectral species concept - an algorithm that 175 clusters pixels from spectral images with a similar spectral signal (referred to 176 as 'spectral species') - has brought attention (Féret & Asner, 2014; Rocchini 177 et al., 2021b). However, in reality, the automatic detection of pixel units 178 sharing a similar spectral signature in a remotely-sensed spectral image does not necessarily match with the actual distribution pattern of a given species 180 but may rather reflect the spatial distribution pattern of a group of species 181 sharing similar biochemical properties (Woodcock and Strahler, 1987). 182

The capacity to identify species is often explained by similarities in spec-183 tral signatures between individuals of the same species, and dissimilarities 184 in spectral signatures between individuals of different species. There is no 185 taxonomic marker in spectroscopy, but individuals of the same species are 186 characterized by a limited set of biophysical and biochemical properties, al-187 lowing differentiation from individuals from other species. Despite signifi-188 cant scientific advances, development of automated retrievals of plant bio-189 chemistry, traits and species identification from satellites across the globe 190 and over time remains aspirational and more work is needed to accomplish 191 this goal, especially for devising a global monitoring of biodiversity change. 192 Here, we aim at introducing the ecological functioning principles of the spec-193 tral species concept and refining its definition by a better linkage with field 194 observations of plant species distribution data (i.e., presence-absence data) 195

available from vegetation surveys. Finally, we will deeply face complexities, challenges and opportunities associated with the use of this concept
for remotely-sensed biodiversity monitoring, including species richness and
evenness (alpha-diversity) as well as composition turnover (beta-diversity).

1.1 A constellation of optical sensors to explore the spectral domain of the Earth surface

The current generation of Earth observing spaceborne and airborne sensors 202 acquires highly-resolved images over a wide range of wavelengths in the solar 203 (optical) and microwave domains of the electromagnetic spectrum (Ustin & 204 Middleton, 2021). The information measured by optical sensors operating in 205 the visible (380-750 nm) and near to shortwave infrared (750-2500 nm) re-206 gions largely corresponds to the region of solar reflectance from the Earth's 207 surface, and are acquired for each individual pixel of an image. Image acqui-208 sition can occur at fine temporal resolutions of days to weeks and in a few 209 cases even at multiple revisits during the day, but data availability and qual-210 ity are strongly dependent on atmospheric conditions. Optical remote sensing 211 systems include those that measure a few discrete spectral bands (multispec-212 tral imaging), such as: (i) the Thematic Mapper on USGS Landsat; (ii) 213 MERIS (Medium Resolution Imaging Spectrometer) on the ESA Envisat; 214 (iii) MODIS (Moderate Resolution Imaging Spectrometer) on the NASA 215 Terra and Aqua platforms; and more recently (iv) the Sentinel-2 satellites of 216 the ESA Copernicus program. These instruments provide open access data 217 describing the Earth surface at frequent intervals and are complemented by 218 commercial satellites that have high spatial resolution like IKONOS, SPOT, 219 Quickbird, WorldView and more recently the Planet constellation. 220

Some spaceborne sensors provide enhanced spectral capabilities and mea-221 sure surface reflectance over hundreds of narrow contiguous spectral bands 222 covering the solar radiation spectrum. These hyperspectral satellites are also 223 called imaging spectrometers. The term hyperspectral emphasizes instru-224 ments measuring a large number of spectral bands while imaging spectrom-225 eter refers to the type of instrument used, i.e., a spectrometer that measures 226 bands across a spectral wavelength interval (spectrum) and produces a 2D ar-227 ray of pixels and spectra. Hyperspectral satellites have demonstrated strong 228 potential for characterizing the chemical and physical structure of the Earth 229 surface, with applications in mineralogy, soil sciences, and vegetation sciences 230 (Plaza et al., 2009). Such images reveal details that improve estimates of 231 key vegetation properties and better discriminate between vegetation types, 232 species or even between genotypes of a given species when the spatial and 233

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spectral resolutions are sufficient to match individuals (Madritch et al., 2014; Blonder et al., 2020). Several satellites hosting imaging spectrometers were 235 recently launched, and a multitude of satellite missions are in preparation 236 (see the review paper by Ustin & Middleton, 2021). These include DESIS 237 and HISUI on the International Space Station (ISS), the free flying platforms 238 PRISMA and EnMAP (expected launch in 2022), along with NASA's and ESA's global monitoring imaging spectrometer missions SBG and CHIME 240 (expected launches in the late 2020s).

Remote sensing tools for monitoring biodiversity 1.2

Species identification from the measurement of the absorptive and reflective 243 characteristics of plants is based on the hypothesis that individuals from the 244 same species share similar biochemical properties, leading to similar spectral 245 characteristics measured at the pixel scale. However, individuals from the 246 same species may also share similar biochemical properties with individuals 247 from another species which may limit our abilities to assign a given spectral 248 signature to a given species. Besides, one needs to also consider phenological 249 patterns, since remote sensing data is generally only acquired during spe-250 cific periods of time (e.g. seasonal), or under specific conditions (e.g. health 251 status), or with specific sensor information (e.g. spatial, spectral and ra-252 diometric resolutions across all wavelengths), all of which may complicate 253 species identification or discrimination. Thus, there are several challenges to 254 identifying a taxonomically identified species with a unique spectral signa-255 ture. Previous research has shown that minor shifts in plant development 256 induced by the environment and its interaction with different plant geno-257 types (i.e., phenotypic variability) may result in co-location between vegeta-258 tive, flowering, fruiting, or senescent stages, each of which expresses different 259 biophysical and biochemical properties. Hence, multiple spectral signatures 260 corresponding to multiple biochemical traits may exist for the same species. 261 Although adding complexity, these problems have solutions. For example, 262 fast growing annual species and facultative annuals (e.g., the invasive Water 263 Hyacinth (*Eichhornia crassipes*) are often found in different growth stages 264 and detection requires several spectral signatures to account for the different 265 growth stages that are later combined for mapping the focal species of con-266 cern (Khanna et al., 2011). Further, spectral discrimination among species is 267 difficult, for example when several species in an ecosystem may share a suite 268 of traits (i.e., strong overlap in the trait space) due to climate constraints and 269 environmental filtering. This happens when trait combinations make individ-270 ual species phenotypically similar (e.g., grass species in grassland habitats, 271 as shown in diversity studies by Gholizadeh et al., 2019, 2020), or when dif-272

ferent combinations of traits result in similar spectral signatures (Kokaly et
al., 2009; Ollinger, 2011). Additional empirical evidence is thus necessary to
contextualize and resolve these questions (Andrew and Ustin, 2008), such as
the data provided by field data collections, experimental studies, and modeling, to achieve full ecological understanding of information from the satellite
data to monitor biodiversity.

²⁷⁹ 1.3 From optical types to the spectral species concept

Just as traditional biodiversity theory focuses on differences between individ-280 uals to assign individuals to different species in order to assess species rich-281 ness, diversity measured with imaging spectrometer data from spaceborne 282 and airborne optical sensors is based on pixels, each with its own spectral 283 information (i.e., reflectance) (Rocchini et al., 2021a). The term spectral 284 signature is more specific than spectral information and usually applied to a 285 specific type of surface (soil, vegetation, water), to a specific material (i.e., 286 traits like chlorophyll concentration), or to a specific level of biological or-287 ganization (genotype, population, species, stand, community, ecosystem). 288 Spectral diversity corresponds to the spatial variation of spectral informa-289 tion: it is tightly related to the notion of (multivariate) variation among 290 species traits (Rocchini et al., 2018) which is the basis of functional diversity 291 in classical ecological theory. In other terms, spectral diversity reflects - at 292 least to a large extent - diversity in community functioning based on the as-293 sortment of functional traits in the community, irrespective of the species/in-294 dividuals that possess these traits (Petchey & Gaston, 2002; Matson et al., 295 2005). Therefore, spectral diversity is conceptually closer to the notion of 296 functional diversity than to the concept of species or taxonomic diversity. 297 For this reason, it is important to keep in mind that the spectral signature 298 of a given pixel unit in a spectral image cannot directly be assigned to a 299 given individual belonging to a specific taxonomic entity (i.e., the spectral 300 species concept). First of all, the pixel size of the image may not match 301 with an actual individual in the field but may contain several individuals be-302 longing to the same species (i.e., a population) or several individuals belong 303 to different species (i.e., a community) such that the same species could be 304 involved in different spectral signatures or 'optical types' (Figure 1). Addi-305 tionally, the same suite of biochemical traits (e.g., foliar nutrient contents) 306 may show extremely high spectral variability depending on the health status 307 of the focal plant species such that different optical types could be assigned 308 to the same species depending on its status: healthy vs. stressed (Figure 309 1, panels f,g,h). Hence, prior to defining the spectral species concept, one 310 needs to identify the different optical types occurring within a given spectral 311

image and then assess how each optical type overlaps with the concept of 312 taxonomic species so that one can assign an optical type to a specific species 313 (i.e., the spectral species concept) or a group of species. To confirm that 314 the spatial distribution pattern of pixels sharing a similar spectral signature 315 (i.e., an optical type) can be directly assigned to the spatial distribution pat-316 tern of individuals from a single species (i.e., the spectral species), one needs 317 ground-truth data on species distribution (i.e., presence-absence data from 318 vegetation surveys) from within the spectral image. 319

Depending on pixel size and size of individuals, pixels can cover multiple 320 individuals or parts of an individual, or stands of one species or mixed stands 321 (Figure 1). In forests, where the spectral species concept was born, a match 322 between spectral and biological species was furthered by the fact that pixels 323 have a higher chance to belong to one species alone (Figure 1, panels a to 324 e, with small pixel size). This does not apply to all situations. Grassland 325 communities, for example, pose a greater challenge because pixels regularly 326 contain several species and optical types can at best be linked to entire 327 communities (Figure 1, panels f to h with any pixel size). These, however, 328 are even less clearly delimited objects than species. 329

³³⁰ 1.4 Spectral species translating spatial distributions of ³³¹ optical types into diversity metrics

Various analytical approaches have been developed to take advantage of the 332 spectral information in the optical domain and investigate different dimen-333 sions of biodiversity (see the review paper by Rocchini et al., 2018). These 334 approaches have demonstrated that it is possible to map and understand 335 functional and taxonomic diversity through space and time, achieved through 336 methodological approaches that differ in their focus, either statistical or 337 process-based. They include data transformation (Rocchini et al., 2017), 338 feature selection and dimensionality reduction (Feilhauer et al., 2011), and 339 machine learning techniques (Kitzes et al., 2021). 340

The number of plant optical types proposed by Ustin & Gamon (2010) is 341 consistent with the Spectral Variation Hypothesis (SVH) (Palmer et al., 2002; 342 Rocchini et al., 2010) which states that ecosystem heterogeneity is associated 343 with high spectral variability. In other words, increased environmental het-344 erogeneity provides more niches for species to co-occur in geographical space 345 with an expected increase in local species richness (Palmer et al., 2002). 346 For optical types, we therefore expect greater diversity where greater spatial 347 environmental heterogeneity occurs. 348

Spectral species - i.e., the number of spectrally distinct classes that ap-

proximate species - are based on the hypothesis that proper processing of 350 images allows discrimination among species, groups of species, or functional 351 groups. Spectral species aim to discretize remotely-sensed information into 352 groups of pixels through unsupervised clustering. The notion of spectral 353 species was successfully applied for mapping tropical biodiversity using air-354 borne high spatial resolution imaging spectroscopy (Féret & Asner, 2014), 355 under the assumption that the majority of pixels did not contain plant mix-356 tures and that individual pixels could be assigned to a species, given that 357 spectral variance among pixels meet statistical criteria and that the average 358 tree crown size approximates the pixel size. Underwood et al. (2007) and 359 Thorp al. (2013) evaluated information content for species mapping from 360 different spatial and spectral resolutions, concluding that high spectral reso-361 lution contributed substantially more information for species mapping than 362 higher spatial resolution. Hence, spectral species and optical types are con-363 ceptually equivalent in their purpose to discriminate among optical entities 364 acquired from an image, but the appropriateness of the terminology may 365 vary with the type of ecosystem and spatial resolution of the sensor. 366

A spectral species is then a set of pixels having similar spectral prop-367 erties that can be used as a proxy for ecologically relevant taxonomic or 368 functional groups, and eventually inventoried to calculate biodiversity met-369 rics like alpha-diversity (e.g., Shannon's H, Simpson's D, Rényi's H) and 370 beta-diversity (e.g., Bray-Curtis dissimilarity or Rao's quadratic entropy) 371 across a landscape ($\mathbf{Box} \mathbf{1}$). Clustering approaches similar to the spectral 372 species framework have been successfully used to map plant species diversity 373 over very different habitats and geographical regions: from African savan-374 nas (Baldeck et al., 2014) to grasslands in the Platte River ecosystem near 375 Wood River, Nebraska, USA (Gholizadeh et al., 2020), and from the Peru-376 vian Andes-Amazon tropical forests (Féret & Asner, 2014) to old-growth, 377 secondary and artificial forests of the Shennongjia National Forest Natural 378 Reserve in China (Zhao et al., 2018). 379

Spectral species combined with functional trait estimation contribute to 380 biodiversity understanding by characterizing the morphological (canopy ar-381 chitecture, gap fraction, etc.) and functional traits, that define the functional 382 role of a species in an ecosystem, e.g. carbon capture strategies by resource 383 acquisition or resource conservation strategies (like drought responses), that 384 then provide a basis to connect spectral species with the taxonomic species 385 that exhibit these trait assemblages. Therefore, spectral species provide a 386 basis to connect with biological species and eventually explore spatial dis-387 tribution of trait assemblages, their evolution in time, and their linkages to 388 environmental or human factors. 389

2 Complexities and challenges

Many of the uncertainties of the spectral species concept apply generally to 391 remotely sensed classified images, which we will not review in detail because 392 they are extensively addressed elsewhere (e.g., Lu & Weng, 2005; Kun et al., 393 2011). The spectral species method of identifying spectral species with dis-394 tinctive traits facilitates identification of rare species that occupy few pixels 395 in an image. Rare and endemic plants by definition do not grow in large 396 colonies of many individuals and generally have restricted distributions. But 397 they are of importance to biodiversity mapping because there are far more 398 rare species than frequent species and they are often concentrated in areas 399 of high biodiversity, especially in hotspots (Griggs, 1940; Medail and Quezel, 400 1997; Ricotta et al., 2010). In terms of biodiversity they may represent 401 species with unique morphological features and genetic richness (Myers et 402 al., 2000; Joppa et al., 2011). Other aspects of uncertainty around the oper-403 ationalization of the spectral species concept are similar to many problems 404 encountered in vegetation classification systems, such as whether rare species 405 are retained in the classified data, whether the classes are realistic in terms 406 of field measured data, and how well the method deals with mixed pixels 407 (multispecies pixels, species with multiple phenotypes). 408

Some of the complications that can make the spectral species concept 409 difficult to handle are shown in the aforementioned Figure 1, where differences 410 in reflection do not necessarily adhere to species boundaries. In fact, several 411 factors determine whether the spectral species concept predicts the current 412 number of species in the area under study. First, more fundamentally it 413 depends on how the taxonomic species are determined, e.g., whether related 414 taxa are characterized by 'splitting' or 'lumping' criteria, leading to more or 415 fewer species. There are also cases of cryptic diversity, i.e., species that are 416 not morphologically or phenotypically distinct, but having a distinct genetic 417 make-up. Thus, the actual number of detected species approximates the true 418 number. Second, even if species are clearly assigned, the spectral species 419 concept requires the measurement of optically active traits, which may not 420 be the traits that differentiate species, or - because of trade-offs (based on the 421 species functional strategy) or selective pressures (resulting from competition, 422 human use and other types of ecological interactions) - they may be not 423 expressed at the time of measurement. Third, community composition and 424 abundance interact with what is detectable – as the sensor primarily measures 425 top of canopy dominant species, which may identify important ecosystem 426 functions but miss much of the diversity. 427

The spectral species concept, as with most remotely sensed measures, is dependent on the temporal dimension to which it is applied. For instance,

measuring the phenological differences among co-occurring plant species through-430 out the season would improve species detection. Including seasonal changes 431 in the analysis may improve detection of understory species when the decid-432 uous overstory is dormant. Multitemporal datasets can form multi-seasonal 433 spectral signatures to account for seasonal changes in a species (e.g., Somers 434 & Asner, 2014) and multitemporal spectral libraries are under development 435 (Dudley et al., 2015). The spectral species concept has shown consistency in 436 identifying taxonomic species and traits from other taxonomic species and 437 their traits. Detecting species in mixed pixels depends on the sensor traits 438 listed above and the magnitude of the differences among the spectral signa-439 tures of the species in the pixel. Species with different functional strategies 440 that are distinguishable in optically active traits will be easier to identify 441 than species with more similar trait assemblies, but which are not optically 442 active. Differences in traits due to leaf types (evergreen or deciduous, needle 443 leaf or broadleaf) create considerable differences in leaf reflectance and timing 444 the date of data acquisition to maximize phenological differences generally 445 always improves the ability to detect sub pixel species. The spectral species 446 concept has yet to be widely tested over the global range of ecosystems and 447 environmental conditions so there is a need to determine its performance and 448 limitations at these scales. 449

2.1 Ecological issues behind spectral species

The spectral species concept was initially developed to allow computation 451 of diversity indices usually computed from species inventories, but with pix-452 els used instead of species. It did not intend to directly estimate 'abso-453 lute' species richness from the spectral clustering, but rather hypothesized 454 that relative species richness and higher-level diversity indices integrating 455 richness and abundance or dissimilarity over space, could be estimated and 456 compared within a limited spatial extent corresponding to an airborne imag-457 ing spectroscopy acquisition. Even in a very unlikely situation of perfect 458 spectral discrimination among species in a remotely sensed acquisition, the 459 predicted number of species may be smaller than the actual species count 460 due to problems related to the potential similarity of traits among multiple 461 species, especially when the growth forms and phenology are similar. The 462 predicted number of species may also be higher than the actual species count 463 if individuals from the same species show phenological shifts. The naming 464 of spectral classes requires matching the taxonomic species information with 465 the spectral information. If the taxonomic naming was based on extreme 466 clumping of groups into fewer taxa or the splitting of taxa into increased 467 number of taxa, the predicted spectral species may differ from the number of 468

⁴⁶⁹ biological taxa identified at a site. Perhaps knowledge of the predicted num⁴⁷⁰ ber of spectral species might result in rethinking the criteria for recognition
⁴⁷¹ and taxonomic revisions, or similarly knowledge of the species may require
⁴⁷² rethinking of how spectroscopy data were analyzed.

While ecological communities have a variety of either dominant, co-dominant 473 and rare species, the latter contribute a lot to the highest diversity measur-474 able on Earth. This is a problem for remote sensing approaches based on 475 clustering, as these algorithms often delete clusters with just a few data 476 points (minimum class size or minimum/maximum number of classes may 477 be selected in the setup), thus there is a tendency to lose rare and endemic 478 species likely fundamental for biological diversity and ecosystem functions. A 479 'continuous surface' analytical method that directly addresses this problem 480 is the widely used 'Multiple Endmember Mixture Model' (MESMA, Roberts 481 et al., 1998), which models each pixel from a range of 'endmembers' (also 482 referred to as 'pure spectra') identified statistically from the convex hull of 483 the data hypervolume. The model can retain classes with small numbers of 484 endmembers depending on user criteria but may have the difficulty in as-485 signing class labels when the variation is continuous (e.g., across a natural 486 continuum or from one habitat type to another, Schmidtlein et al., 2007). 487

Environmental heterogeneity may also affect the accuracy of the spectral 488 species methodology (Schmidtlein & Fassnacht, 2017). Disturbances may 489 increase or decrease biodiversity, with a net zero effect on heterogeneity. 490 There are several examples (e.g., forest gap openings, fire spread, and urban 491 development) that can increase heterogeneity without an increase overall bi-492 ological diversity. Spectral species should detect these changes as differences 493 in the assemblage and range of traits being measured, although there is un-494 certainty about the least detectable magnitude of change. Therefore, we 495 expect remotely sensed spectral species to provide a first-order exploratory 496 tool to, at least detect areas that are suspected of hosting a high number 497 of species or, alternatively, identify areas that should host larger numbers of 498 species (based on independent criteria) but don't. From this point of view, 499 in-situ data would greatly improve the classification, validate its accuracy, 500 and provide a basis to investigate and identify sites that do not match a 501 priori expectations (Foody et al., 2016). 502

The R package (biodivMapR) dedicated to applications of the spectral species algorithm (Féret and de Boissieu, 2020) on raster data requires the number of spectral species be defined by the user. The method is not intended to estimate the absolute number of species from remote sensing, as this problem is highly scale and context dependent as explained earlier. The definition of the optimal number of clusters in a dataset is a specific problem

that the spectral species framework was also not intended to solve. Adding 510 a method to determine the number of clusters corresponding to species rich-511 ness over a full dataset would definitely contribute to the generalization of the 512 spectral species framework over various ecosystems. The number of species 513 included in a remote sensing acquisition is not generally known a priori for 514 most of the Earth's ecosystems. However, there are statistical approaches 515 to estimate the number with enough accuracy that realistic values could be 516 produced (Chang and Du, 2004; Gholizadeh et al., 2020), as well as deeply 517 rooted ecological principles based on species dispersal, biogeography, land-518 scape ecology (Rocchini et al., 2021a). 519

There are questions on whether this method will continue to be successful 520 at the spatial scales of satellites, where multiple species within pixels cause 521 spectral mixing, and under conditions of sparse vegetation where mixed pix-522 els can include plants, soils, plant litter, geological minerals, water, ice and 523 man-made materials. This will be resolved with the use and applications 524 of the new generation of hyperspectral satellites. The various spaceborne 525 imaging spectrometers (see the review paper by Ustin & Middleton, 2021) 526 that are available now (DESIS, EnMAP, HISUI, PRISMA) or in the next few 527 years (e.g., CHARM, EMIT, SBG) all have 30m pixels. These imaging spec-528 trometers will provide ample opportunities to examine how well the spectral 529 species concept scales up and across ecosystems. 530

In addition to passive optical sensors, reviewed in this manuscript, LiDAR 531 and radar active sensors might complement observations with the capacity 532 to penetrate the uppermost layer of vegetation and provide information on 533 canopy structure, while being less sensitive to atmospheric perturbations 534 than passive optical sensors, Asner et al., 2008; Bergen et al., 2009; Zhao et 535 al., 2018; Mulatu et al., 2019; Simonson et al., 2021; Lenoir et al., in press). 536 For instance, the GEDI LiDAR mission on the ISS is measuring at high 537 vertical and spatial resolutions the distribution and height of global woody 538 vegetation to provide structural information, including understory data, for 539 biodiversity estimates. Besides, the NISAR and Biomass radar missions 540 (launches in 2023) will monitor global patterns of biomass, disturbances, 541 and impacts on biodiversity. Together, advances in the spatial, spectral and 542 temporal dimensions of imagery offer immense data streams that can be har-543 nessed to better understand the processes of biological functioning, and to 544 systematically map and monitor ecosystem changes from local to regional to 545 global scale. 546

In this manuscript we focused on the need to model biodiversity from hy-548 perspectral remote sensing and field data to calibrate such models. A huge 549 amount of field (free) data is actually available (e.g., sPlotOpen, GBIF), but 550 creativity for mapping species diversity from hyperspectral imagery should be 551 reinforced. From this point of view, the spectral species concept is expected 552 to make a major contribution in mapping and analysis of hyperspectral satel-553 lite data to produce remote sensing based essential biodiversity variables 554 (RS-EBVs, see Skidmore et al., 2015, 2021). Even if the predicted spectral 555 species do not precisely identify the number and identity of taxonomic species 556 in an image, it will provide a first-order exploratory tool to detect areas of 557 low to high species diversity. Such approaches will require a better under-558 standing of the range of conditions under which the spectral species concept 559 can operate. In section 2 we highlight some limitations for applications of 560 the concept, but many of these limitations should be seen as new opportu-561 nities for research. These include, among others, identification of additional 562 spectral characteristics that can be captured and what additional traits, and 563 thus more species will be detected. We are aware that very small absorption 564 features are present in plant spectra and are statistically detectable but we 565 do not know the physical basis of the biochemical/biophysical material that 566 is absorbing this energy, highlighting an important avenue of research for the 567 future. There is an ever-growing number of new traits reported from imaging 568 spectroscopy research, ranging from phenolic and isoprene compounds, non-569 structural carbohydrates, fiber content (Serbin et al., 2014; Singh et al., 2015; Ely et al., 2019), essential nutrients including potassium, phosphorous, and calcium (Asner et al., 2015; Elv et al., 2019) to RuBP carboxylation (Vcmax) and regeneration (Jmax) (Serbin et al., 2012, 2014; Rogers et al., 2017; Wu et al., 2019). Such information is retrieved from measurements enabled by today's imaging spectrometers, which typically have spectral resolutions of 10 nanometer (nm) spectral bands cross the solar spectrum for a total of around 200-250 bands, some have 3-5nm in the visible-near-infrared wavelengths and 5-10nm in the shortwave infrared region. New satellite sensors under development (Flora on the Flex platform and GeoCarb) are designed to detect chlorophyll fluorescence with 1-3nm narrow bands in the wavelength regions where chlorophyll fluoresces near the oxygen bands.

The spectral species concept is based on the principle that the variability in a spectral data cube is sufficient to identify a suite of traits in a pixel. In these cases, pixels are composed of multiple species with an assortment of different traits. The spectral species concept has been broadened to that of "spectranomics", which is "an approach to conceptually and geographically

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link plant canopy species and their functional traits to their spectral-optical properties" (Asner & Martin, 2016). Broadening "spectranomics" has benefit for biodiversity assessment across different organizational levels, another area in which the spectral species concept could contribute. 590

While differences in traits identify different species, they can also describe 592 patterns at different levels of ecological organization: habitat, community, 593 ecosystem. In Ustin & Gamon (2010), the concept of "plant optical types" is based on their spectral attributes, without any direct reference to species 595 or traits. By relating these optical types to specific traits (e.g. leaf and 596 canopy resources allocated to productivity) these optical types acquire the definition of plant functional types. Just as actual plant species change, 598 the spectral species concept must be dynamic and change. As the range 599 of applications increase, the range of spectral patterns will increase (e.g., 600 including phenological events like mass flowering or senescent states) (Poyry et al., 2018). 602

Understanding the interactions between biodiversity and ecological/envi-603 ronmental drivers is difficult (Kreft & Jetz, 2007). From this point of view, 604 collecting exploratory remote sensing data on environmental heterogeneity 605 across large geographical extents is relatively simple and combining that in-606 formation with changing patterns of functional traits could improve species 607 identification, mapping and monitoring of potential diversity hotspots (Skid-608 more et al., 2015; Asner et al., 2017). 609

As biodiversity research expands to global ecosystems, there are many 610 questions unresolved. Both ecological sciences and Earth observation tech-611 nologies are still in maturing phases of development. In this paper, we clari-612 fied the links between the spectral species concept, optical types and optical 613 traits, and their analogy with ecological dimensions including species, func-614 tional types and functional traits, as well as the mechanistic link between 615 biophysical properties of vegetation and what is usually expressed as spectral 616 signatures, corresponding to species. The relevance of discrete approaches, 617 and their complementarity with continuous approaches was also highlighted. 618

As stressed in this paper, important advances have been made in under-619 standing how the spectral signature relates to biodiversity and where there 620 is untapped potential to further disentangle these connections. While one 621 should not overestimate the capacity of remote sensing to directly estimate 622 biodiversity from space (Skidmore et al., 2021), remote sensing captures pat-623 terns of reflected or emitted electromagnetic radiation that are driven by bio-624 physical and biochemical properties of vegetation (i.e., patterns dependent 625 on the optical properties of plants that are the phenotypic expressions of their 626 fitness strategies). There are a wide range of scientific disciplines that include 627

all ecological subdisciplines from the population to the global biosphere, bi-628 ology, soils, hydrology, evolution and phylogenetics, optics, mathematics, 629 statistics and informatics and engineering, that are necessary to understand, 630 interpret, refine and improve remote sensing research. To actually develop 631 and verify methods to identify and monitor global patterns of biodiversity, it 632 will truly take "more than a village" but instead an engaged and committed 633 international contingent of scientists, social scientists, citizen scientists, en-634 gineers, policy makers, land owners, farmers, and more, to shorten the long 635 road ahead. 636

4 Authors' contributions

DR, MJS, SLU, JBF and JL led the writing of this manuscript. DR, JBF, GPA, MD, MJS and SS analyzed the data providing the output for the figures of this manuscript. SS conceived and crafted Figure 1. All authors contributed critically to the development of the concept of this manuscript and to the writing of the draft, and they gave final approval for publication.

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657 References

Andrew, M.E., Ustin, S.L. (2008). The role of environmental context in map ping invasive plants with hyperspectral image data. Remote Sensing of
 Environment, 112, 4301-4317.

Asner, G.P., Knapp, D.E., Kennedy-Bowdoin, T., Jones, M.O., Martin, R.E.,
Boardman, J. & Hughes, R.F. (2008a) Invasive species detection in Hawaiian rainforests using airborne imaging spectroscopy and LiDAR. Remote
Sensing of Environment, 112, 1942-1955.

Asner, G.P. & Martin, R.E. (2008) Spectral and chemical analysis of tropical
 forests: scaling from leaf to canopy levels. Remote Sensing of Environment,
 112, 3958-3970.

Asner, G.P., Martin, R.E. (2016) Spectranomics: Emerging science and con servation opportunities at the interface of biodiversity and remote sensing.
 Global Ecology and Conservation, 8, 212-219.

Asner, G.P., Martin, R.E., Anderson, C.B., Knapp, D.E. (2015) Quantifying
forest canopy traits: Imaging spectroscopy versus field survey. Remote
Sensing of Environment, 158, 15-27.

Asner, G.P., Martin, R.E., Knapp, D.E., Tupayachi, R., Anderson, C.B.,
Sinca, F., Vaughn, N.R. & Llactayo, W. (2017) Airborne laser-guided
imaging spectroscopy to map forest trait diversity and guide conservation.
Science, 355, 385-389.

Baldeck, C.A., Colgan, M.S., Féret, J.B., Levick, S.R., Martin, R.E. & Asner,
G.P. (2014) Landsape-scale variation in plant community composition of an
African savanna from airborne species mapping. Ecological Applications,
24, 84-93.

Baret, F., Vanderbilt, V.C., Steven, M.D. & Jacquemoud, S. (1994) Use of
spectral analogy to evaluate canopy reflectance sensitivity to leaf optical
properties. Remote Sensing of Environment, 48, 253-260.

Bergen, K.M., Goetz, S.J., Dubayah, R.O., Henebry, G.M. et al. (2009)
Remote sensing of vegetation 3-D structure for biodiversity and habitat:
Review and implications for lidar and radar spaceborne missions. Journal
of Geophysical Research: Biogeosciences, 114, G00E06.

Blonder, B., Graae, B.J., Greer, B., et al. (2020). Remote sensing of ploidy
level in quaking aspen (*Populus tremuloides* Michx.). Journal of Ecology,
108, 175-188.

Cawse-Nicholson, K., Townsend, P.A., Schimel, D., Assiri, A.M., Blake, P.L.
et al. (2021). NASA's surface biology and geology designated observable:
a perspective on surface imaging algorithms. Remote Sensing of Environment, 257, 112349.

⁶⁹⁶ Chang, C.I., Du, Q. (2004) Estimation of number of spectrally distinct signal
 ⁶⁹⁷ sources in hyperspectral imagery. IEEE Transactions on Geoscience and
 ⁶⁹⁸ Remote Sensing 42: 608-619.

⁶⁹⁹ Dudley, K.L., Dennison, P.E., Roth, K.L., Roberts, D.A., Coates, A.R.
⁷⁰⁰ (2015). A multi-temporal spectral library approach for mapping vegeta⁷⁰¹ tion species across spatial and temporal phenological gradients. Remote
⁷⁰² Sensing of Environment, 167, 121-134.

Ely, K.S., Burnett, A.C., Lieberman-Cribbin, W., Serbin, S.P. & Rogers, A.
(2019) Spectroscopy can predict key leaf traits associated with source-sink
balance and carbon - nitrogen status. Journal of Experimental Botany, 70,
1789-1799.

Ewald, M., Skowronek, S., Aerts, R., Dolos, K., Lenoir, J., Nicolas, M.,
Warrie, J., et al. (2018). Analyzing remotely sensed structural and chemical
canopy traits of a forest invaded by *Prunus serotina* over multiple spatial
scales. Biological Invasions, 20, 2257-2271.

Fassnacht, F.E., Latifi, H., Stereńczak, K., Modzelewska, A., Lefsky, M.,
Waser, L.T., Straub, C. & Ghosh, A. (2016). Review of studies on tree
species classification from remotely sensed data. Remote Sensing of Environment, 186, 64-87.

Feilhauer, H., Faude, U. & Schmidtlein, S. (2011) Combining Isomap ordination and imaging spectroscopy to map continuous floristic gradients in a heterogeneous landscape. Remote Sensing of Environment, 115, 2513-2524.

Feilhauer, H., Somers, B., van der Linden, S. (2017) Optical trait indicators
for remote sensing of plant species composition: Predictive power and
seasonal variability. Ecological Indicators, 73, 825-833.

Féret, J.-B. & Asner, G.P. (2014) Mapping tropical forest canopy diversity
using high-fidelity imaging spectroscopy. Ecological Applications, 24, 12891296.

Féret, J.-B. & de Boissieu, F. (2020) biodivMapR: An R package for α - and β -diversity mapping using remotely sensed images. Methods in Ecology and Evolution, 11, 64- 70. Féret, J.-B., François, C., Asner, G.P., Gitelson, A.A., Martin, R.E., Bidel,
L.P.R., Ustin, S.L. & le Maire, G., Jacquemoud, S. (2008) PROSPECT-4
and 5: Advances in the leaf optical properties model separating photosynthetic pigments. Remote Sensing of Environment, 112, 3030-3043.

Foody, G.M., Pal, M., Rocchini, D., Garzon-Lopez, C.X. Bastin, L. (2016)
The sensitivity of mapping methods to reference data quality: Training
supervised image classifications with imperfect reference data. ISPRS International Journal of Geo-Information 5: 199.

⁷³⁵ Gholizadeh, H., Gamon, J.A., Helzer, C. J. & Cavender-Bares, J. (2020) ⁷³⁶ Multi-temporal assessment of grassland α - and β -diversity using hyper-⁷³⁷ spectral imaging. Ecological Applications 30: e02145.

Gholizadeh, H., Gamon, J.A., Townsend, P.A., Zygielbaum, A.I., Helzer,
C.J., Hmimina, G.Y., Yu, R., Moore, R.M., Schweiger, A.K., CavenderBares, J. (2019) Detecting prairie biodiversity with airborne remote sensing. Remote Sensing of Environment 221: 38-49.

Griggs, R.F. (1940) The ecology of rare plants. Bulletin of the Torrey Botanical Club, 67, 575-594.

Khanna, S., Santos, M.J., Ustin, S.L. & Haverkamp, P.J. (2011) An integrated approach to a biophysiologically based classification of floating
aquatic macrophytes. International Journal of Remote Sensing, 32, 10671094.

Kitzes, J., Blake, R., Bombaci, S., Chapman, M., Duran, S. M. et al. (2021)
Expanding NEON biodiversity surveys with new instrumentation and machine learning approaches. Ecosphere, 12, e03795.

Kokaly, R.F., Asner, G.P., Ollinger, S.V., Martin, M.E. & Wessman, C.A.
(2009) Characterizing canopy biochemistry from imaging spectroscopy and its application to ecosystem studies. Remote Sensing of Environment, 113, 78-91.

Kreft, H. & Jetz, W. (2007) Global patterns and determinants of vascular
plant diversity. Proceedings of the National Academy of Sciences, 104,
5925-5930.

Kun, J., Li, Q., Tian, Y.-C. & Wu, B.-F. (2011) A review of classification
method of remote sensing imagery. Spectroscopy and Spectral Analysis,
31, 2618-1623.

⁷⁶¹ Kwok, R. (2018). Ecology's remote-sensing revolution. Nature, 556, 137-138.

Jacquemoud, S. & Ustin, S.L. (2019) Leaf Optical Traits. Cambridge Uni versity Press, Cambridge, UK.

Jacquemoud, S., Ustin, S.L., Verdebout, J., Schmuck, G., Andreoli, G. &
Hosgood, B. (1996) Estimating leaf biochemistry using the PROSPECT
leaf optical properties model. Remote Sensing of Environment, 56, 194-202.

Joppa, L.N., Roberts, D.L., Myers, N., Pimm, S.L. (2011) Biodiversity
hotspots house most undiscovered plant species. Proceedings of the National Academy of Sciences of the United State of America, 108, 1317113176.

Lenoir, J., Gril, E., Durrieu, S., Horen, H., Laslier, M., Lembrechts, J. et al.
(in press). Unveil the unseen: Using LiDAR to capture time-lag dynamics
in the herbaceous layer of European temperate forests. Journal of Ecology.
doi: https://doi.org/10.1111/1365-2745.13837

Lu, D. & Weng, Q. (2005) A survey of image classification methods and
techniques for improving classification performance. International Journal
of Remote Sensing, 28, 823-870.

Luque, S., Pettorelli, N., Vihervaara, P. & Wegmann, M. (2018). Improving
biodiversity monitoring using satellite remote sensing to provide solutions
towards the 2020 conservation targets. Methods in Ecology and Evolution,
9, 1784-1786.

Madritch, M.D., Kingdon, C.D., Singh, A., Mock, K.E., Lindroth, R.L.,
Townsend, P.A. (2014) Imaging spectroscopy links aspen genotype with
below-ground processes at landscape scales. Philosophical Transactions of
the Royal Society B., 369, 20130194.

Matson, N.W.H., Mouillot, D., Lee, W.G. & Wilson, J.B. (2005) Functional
richness, functional evenness and functional divergence: the primary components of functional diversity. Oikos, 111, 112-118.

Medail, F. & Quezel, P. (1997). Hot-spots analysis for conservation of plant
biodiversity in the Mediterranean Basin. Annals of the Missouri Botanical
Garden, 84, 112-127.

Myers, N., Mittermeier, R.A., Mittermeier, C.G., de Fonseca, G.A.B. & Kent,
J. (2000) Biodiversity hotspots for conservation priorities. Nature, 403,
853-858.

Mulatu, K.A., Decuyper, M., Brede, B., Kooistra, L., Reiche, J., Mora, B. &
Herold, M. (2019) Linking Terrestrial LiDAR Scanner and Conventional
Forest Structure Measurements with Multi-Modal Satellite Data. Forests,
10, 291.

⁷⁹⁹ Ollinger, S.V. (2011) Sources of variability in canopy reflectance and the ⁸⁰⁰ convergent properties of plants. New Phytologist, 189, 375-394.

Palmer, M.W., Earls, P.G., Hoagland, B.W., White, P.S. & Wohlgemuth, T.
(2002) Quantitative tools for perfecting species lists. Environmetrics, 13, 121-137.

Petchey, O.L. Gaston, K.J. (2002) Functional diversity (FD), species richness
and community composition. Ecology Letters, 5, 402-411.

Plaza, A., Benediktsson, J.A., Boardman, J.W., Brazile, J. et al. (2009).
Recent advances in techniques for hyperspectral image processing. Remote
Sensing of Environment, 113, 110-122.

Pollock, L.J., O'Connor, L.M.J., Mokany, K., Rosauer, D.F., Talluto, M.V.
& Thuiller, M. (2020) Protecting biodiversity (in all its complexity): new models and methods. Trends in Ecology & Evolution, 35, 1119-1128.

Poyry, J., Bottcher, K., Fronzek, S., Gobron, N., Leinonen, R., Metsamaki, S. & Virkkala, R. (2018) Predictive power of remote sensing versus temperature-derived variables in modelling phenology of herbivorous insects. Remote Sensing in Ecology and Conservation, 4: 113-126.

Randin, C.F., M.B Ashcroft, J. Boliger, J. Cavender-Bares, N.C. Coops, et al.
(2020) Modeling biodiversity in the Anthropocene using remote sensing in
species distribution models. Remote Sensing of Environment, 239, 111626.

Ricotta, C., Godefroid, S., Rocchini, D. (2010). Patterns of native and exotic
species richness in the urban flora of Brussels: rejecting the 'rich get richer'
model. Biological Invasions, 12, 233-240.

Roberts, D., Gardner, M., Church, R., Ustin, S., Scheer, G. & Green, R.
(1998) Mapping chaparral in the Santa Monica mountains using multiple
endmember spectral mixture models. Remote Sensing of Environment, 65,
267-279.

Rocchini, D., Balkenhol, N., Carter, G.A., Foody, G.M. Gillespie, T.W., He,
K.S., Kark, S., Levin, N., Lucas, K., Luoto, M., Nagendra, H., Oldeland,
J., Ricotta, C., Southworth, J. & Neteler, M. (2010). Remotely sensed

spectral heterogeneity as a proxy of species diversity: recent advances and open challenges. Ecological Informatics, 5, 318-329.

Rocchini, D., Marcantonio, M., Da Re, D., Bacaro, G., Feoli, E., Foody, 831 G.M., Furrer, R., Harrigan, R.J., Kleijn, D., Iannacito, M., Lenoir, J., 832 Lin, M., Malavasi, M., Marchetto, E., Meyer, R.S., Moudry, V., Schneider, 833 F.D., Simova, P., Thornhill, A.H, Thouverai, E., Vicario, S., Wayne, R.K. 834 & Ricotta, C. (2021). From zero to infinity: minimum to maximum di-835 versity of the planet by spatio-parametric Rao's quadratic entropy. Global 836 Ecology and Biogeography, 30, 1153-1162. 837

Rocchini, D., Marcantonio, M., Ricotta, C. (2017). Measuring Rao's Q di-838 versity index from remote sensing: an open source solution. Ecological 839 Indicators, 72, 234-238. 840

Rocchini, D., Luque, S., Pettorelli, N., Bastin, L., Doktor, D., Faedi, N., 841 Feilhauer, H., Féret, J.-B., Foody, G.M., Gavish, Y., Godinho, S., Kunin, 842 W.E., Lausch, A., Leitão, P.J., Marcantonio, M., Neteler, M., Ricotta, 843 C., Schmidtlein, S., Vihervaara, P., Wegmann, M., Nagendra, H. (2018). 844 Measuring β -diversity by remote sensing: a challenge for biodiversity mon-845 itoring. Methods in Ecology and Evolution, 9, 1787-1798. 846

Rocchini, D., Salvatori, N., Beierkuhnlein, C., de Boissieu, F., Förster, M., Garzon-Lopez, C.X., Gillespie, T.W., Hauffe, H.C., He, K.S., Kleinschmit, 848 B., Lenoir, J., Nagendra, H., Malavasi, M., Moudrý, V., Payne, D., Šímová, 849 P., Torresani, M., Wegmann, M., Féret, J.B. (2021). From local spectral 850 species to global spectral communities: a benchmark for ecosystem diver-851 sity estimate by remote sensing. Ecological Informatics, 61, 101195. 852

Rogers, A., Medlyn, B.E., Dukes, J.S., Bonan, G., von Caemmerer, S. et al. 853 (2017) A roadmap for improving the representation of photosynthesis in 854 Earth system models. New Phytologist, 213, 22-42. 855

Schaaf, A.N., Dennison, P.E., Fryer, G.K., Roth, K.L., Roberts, D.A. (2013) 856 Mapping plant functional types at multiple spatial resolutions using imag-857 ing spectrometry data. GIScience & Remote Sensing, 48, 324-344. 858

Schmidtlein, S. & Fassnacht, F.E. (2017) The spectral variability hypothesis 859 does not hold across landscapes. Remote Sensing of Environment, 192, 860 114-125. 861

Schmidtlein, S., Zimmermann, P., Schupferling, R. & Weiss, C. (2007) Map-862 ping the floristic continuum: Ordination space position estimated from 863 imaging spectroscopy. Journal of Vegetation Science, 18, 131-140. 864

847

829

Schweiger, A.K., Laliberté, E. (2022). Plant beta-diversity across biomes
 captured by imaging spectroscopy. Nature Communications, 13, 2767.

Schweiger, A.K., Schultz, M., Risch, A.C., Kneubuhler, M., Haller, R.,
Schaepman, M.E. (2017). How to predict plant functional types using
imaging spectroscopy: linking vegetation community traits, plant functional types and spectral response. Methods in Ecology and Evolution, 8,
86-95.

Serbin, S.P., Dillaway, D.N., Kruger, E.L., Townsend, P.A. (2012) Leaf optical properties reflect variation in photosynthetic metabolism and its sensitivity to temperature. Journal of Experimental Botany, 63, 489-502.

Serbin, S.P., Singh, A., McNeil, B.E., Kingdon, C.C. & Townsend, P.A.
(2014) Spectroscopic determination of leaf morphological and biochemical
traits for northern temperate and boreal tree species. Ecological Applications, 24, 1651-1669.

Skidmore, A.K., Pettorelli, N., Coops, N.C., Geller, G.N., Hansen, M., Lucas,
R., Mücher, C.A., O'Connor, B., Paganini, M., Pereira, H.M., Schaepman,
M.E., Turner, W., Wang, T., Wegmann, M. (2015) Agree on biodiversity
metrics to track from space. Nature, 523, 403-405.

Skidmore, A.K., Coops, N.C., Neinavaz, E., Ali1, A., et al. (2021) Priority list
of biodiversity metrics to observe from space. Nature Ecology & Evolution,
5, 896-906.

Skowronek, S., Ewald, M., Isermann, M., Van De Kerchove, R., Lenoir, J.,
et al. (2017) Mapping an invasive bryophyte species using hyperspectral
remote sensing data. Biological Invasions, 19, 239-254.

Simonson, W. D., Allen, H. D. & Coomes, D. A. (2012) Use of an airborne
lidar system to model plant species composition and diversity of Mediterranean oak forests. Conservation Biology, 26, 840-850.

Singh, A., Serbin, S.P., McNeil, B.E., Kingdon, C.C., Townsend, P.A. (2015)
Imaging spectroscopy algorithms for mapping canopy foliar chemical and
morphological traits and their uncertainties. Ecological Applications, 35,
2180-2197.

Somers, B. & Asner, G.P. (2014) Tree species mapping in tropical forests
using multi-temporal imaging spectroscopy: wavelength adaptive spectral
mixture analysis. International Journal of Applied Earth Observation and
Geoinformation, 31, 57-66.

Torresani, M., Feilhauer, H., Rocchini, D., Féret, J.-B., Zebisch, M., Tonon,
G. (2021). Which optical traits enable an estimation of tree species diversity based on the Spectral Variation Hypothesis? Applied Vegetation
Science, 24: e12567.

Thorp, K.R., French, A.N. & Rango, A. (2013) Effect of image spatial and
spectral characteristics on mapping semi-arid rangeland vegetation using
multiple endmember spectral mixture analysis (MESMA). Remote Sensing
of Environment, 132, 120-130.

⁹⁰⁸ Underwood, E.C., Ustin, S.L. & Ramirez, C.M. (2007) A comparison of
⁹⁰⁹ spatial and spectral image resolution for mapping invasive plants in coastal
⁹¹⁰ California. Environmental Management, 39, 63-83.

⁹¹¹ Ustin, S.L. & Middleton, E.M. (2021) Current and near-term advances in
⁹¹² Earth observation for ecological applications. Ecological Processes, 1.

⁹¹³ Ustin, S.L. & Gamon, J.A. (2010) Remote sensing of plant functional types.
⁹¹⁴ New Phytologist, 186, 795-816.

⁹¹⁵ Woodcock, C.E. & Strahler, A.H. (1987). The factor of scale in remote sens-⁹¹⁶ ing. Remote sensing of Environment, 21, 311-332.

⁹¹⁷ Wu, J., Rogers, A., Albert, L.P., Ely, K., Prohaska, N. et al. (2019) Leaf re⁹¹⁸ flectance spectroscopy captures variation in carboxylation capacity across
⁹¹⁹ species, canopy environment and leaf age in lowland moist tropical forests.
⁹²⁰ New Phytologist, 224, 663-674.

⁹²¹ Zhao, Y., Zeng, Y., Zheng, Z., Dong, W., Zhao, D., Wu, B. Zhao, Q. (2018)
⁹²² Forest species diversity mapping using airborne LiDAR and hyperspectral
⁹²³ data in a subtropical forest in China. Remote Sensing of Environment, 213,
⁹²⁴ 104-114.

Figures and boxes

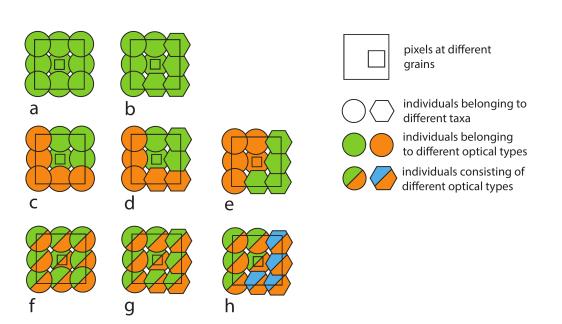


Figure 1: Some potential scenarios that can happen in optical remote sensing of vegetation canopies. The graph shows sources of variation in the relationship between species and optical types. Plants of different species might belong to different optical types, but many other situations can also be found. Optical types can be related to information of interest (e.g., species or plant traits) or to irrelevant pattern (e.g., shadows, depending on the research question). Scenario (a) represents a stand with individuals of only one single species, with a similar reflectance. In scenario (b) individuals of two species have a similar reflectance; hence they would be grouped in the same spectral species. This is further complicated once mixing individuals belonging to the same taxon but to different optical types (c) or individuals of multiple species belonging to different optical types that do not follow the species boundaries (d). What many would hope for is that plants of different species belong to different optical types, which may happen (e). Finally, the same plant individual can consist of different optical types showing different spectral properties in e.g. young versus old leaves, shadow and light, or differences in health conditions. This intra-individual mixing property will be related to all of the previous cases (f-h). Note that a stand or individual can pass through several of these scenarios in time (intra and interannual variability).

Box 1 - The spectral species algorithm at work

The spectral species concept is grounded in an algorithm which is now ready available under a free and open source R package in GitHub (https:// github.com/jbferet/biodivMapR) named biodivMapR (Féret and de Boissieu, 2020), which is able to produce α - and β -diversity maps starting from the detection of spectral species based on the optical properties of vegetation in the field.

Starting from a multi- or hyper-spectral image, a spectral transformation 934 like a Principal Component Analysis (PCA) is performed to reduce the di-935 mensionality for further calculations (Figure 2). Based on those principal 936 components explaining most of the variance, k-means clustering is applied to 937 a random set of pixels in order to detect group of pixel with the same spectral 938 reflectance, possibly related to single groups of individuals/canopies in the 939 field that are sharing similar traits and thus that are likely to be phyloge-940 netically related (spectral species concept). The detection of spectral species 941 is applied backward to the whole map defining a membership probability of 942 each pixel to a certain spectral species, as based on its spectral euclidean 943 distance from the centroid of the previously defined clusters. A single fi-944 nal spectral species map can be attained by using the maximum probability 945 for each pixel to attain to a certain spectral species. The spectral species 946 map is further divided into elementary spatial units with a higher pixel di-947 mension (hereafter simply units), in which calculation of α - and β -diversity 948 can be performed, leading to crucial information on both local diversity and 949 turnover. In order to attain α -diversity, Shannon's H' is calculated for each 950 unit, while for β -diversity calculation Bray-Curtis dissimilarity is considered 951 among all the possible pairs of units. The α -diversity map can be shown by 952 directly taking into account Shannon's H' for each unit and reporting it in 953 a final resampled map. Rather, the β -diversity map needs a further step to 954 pass from a distance matrix to a 2D spatial representation, namely the ap-955 plication of Non metric Multidimensional Scaling (NMDS), by deciding the 956 final number of reduced dimensions (in this case three in order to compose 957 an RGB image). The final result will be a unitless β -diversity map in which 958 different colours represent differences among species communities in space. 959

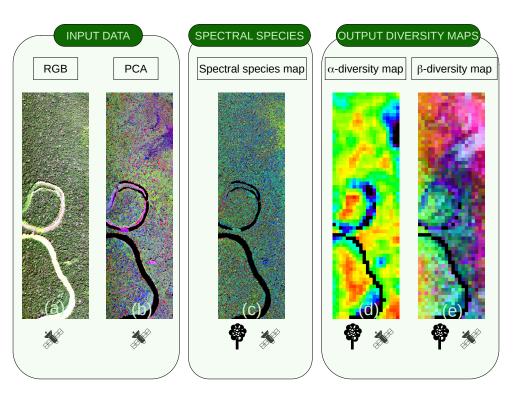


Figure 2: Box 1 Figure - The spectral species algorithm phases. The original image was acquired with the CAO AToMS imaging spectrometer during an airborne campaign over the CICRA experimental site (Amazonian Peru) (https://www.amazonconservation.org/about/ mission-vision/cicra-station/). (a) corresponds to the RGB representation of an imaging spectroscopy subset. A standardized PCA is applied on (a) and a reduced set of components is selected (b), to maximize signal corresponding to biological patterns on forested areas and discard noisy components. Spectral species are defined for each pixel by applying an unsupervised k-means clustering on the spectral space defined by selected components (c). In this phase, a field survey recognition based on in-situ data is crucial to define the number of singular spectral signatures (spectral species) expected. The spectral species map is divided into elementary spatial units and the spectral species inventory is performed for each spatial unit, by further calculating Shannon's H and Bray-Curtis metrics to derive alpha- ((d), ranging here from minima to maxima from black to blue, green and red) and betadiversity ((e), in which colours represent differences among spectral species) maps, respectively.